

ORIGINAL ARTICLE

Relevance of drinking water as a source of human exposure to bisphenol A

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A comprehensive search of studies describing bisphenol A (BPA) concentrations in drinking water and source waters (i.e., surface water and groundwater) was conducted to evaluate the relevance of drinking water as a source of human exposure and risk. Data from 65 papers were evaluated from North America (31), Europe (17), and Asia (17). The fraction of drinking water measurements reported as less than the detection limit is high; 95%, 48%, and 41%, for North America, Europe, and Asia, respectively. The maximum quantified (in excess of the detection limit) BPA concentrations from North America, Europe, and Asia are 0.099 $\mu\text{g}/\text{l}$, 0.014 $\mu\text{g}/\text{l}$, and 0.317 $\mu\text{g}/\text{l}$. The highest quantified median and 95th percentile concentrations of BPA in Asian drinking water are 0.026 $\mu\text{g}/\text{l}$ and 0.19 $\mu\text{g}/\text{l}$, while high detection limits restricted the determination of representative median and 95th percentile concentrations in North America and Europe. BPA in drinking water represents a minor component of overall human exposure, and compared with the lowest available oral toxicity benchmark of 16 $\mu\text{g}/\text{kg-bw}/\text{day}$ (includes an uncertainty factor of 300) gives margins of safety > 1100. Human biomonitoring data indicate that ingestion of drinking water represents <2.8% of the total intake of BPA.

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INTRODUCTION

Bisphenol A (BPA, 4,4'-isopropylidene diphenol, CAS Registry No. 80-05-7) is a commercially important industrial chemical with an estimated worldwide production capacity of approximately 5.2 million metric tonnes in 2008.¹ BPA is primarily used as an intermediate in the production of polycarbonate plastic and epoxy and other specialty resins.^{2–4} Major end-use applications for polycarbonate include glazing and sheeting, electrical and electronic goods, electronic storage media, and household equipment, including bottles, utensils and containers. Epoxy resins are used for protective coatings for architectural structures, marine and car coatings, container coatings, and printed circuit boards. BPA is also used in the production of phenoplast, phenolic and unsaturated polyester resins, polyvinylchloride, and thermal paper. The presence of BPA in the environment and consumer products has been the subject of public and regulatory attention, primarily due to concerns about its weak endocrine activity. The environmental fate and ecotoxicological properties of BPA have been extensively evaluated^{5–8} and a number of risk assessments have been conducted by regulatory authorities around the world.^{3,9,10}

Small amounts of BPA may enter the environment from production and processing facilities, which often discharge to sewage treatment plants.^{3,6} Once introduced to the environment, BPA primarily partitions to the aquatic compartment.⁵ Extensive monitoring of BPA in various environmental media has been conducted over the last 10 years. Klecka et al.¹¹ recently published the results of a statistical analysis of environmental concentrations in North America and Europe. Median BPA concentrations for

fresh surface waters for North America and Europe were 0.081 $\mu\text{g}/\text{l}$ and 0.01 $\mu\text{g}/\text{l}$, while 95th percentiles were 0.47 $\mu\text{g}/\text{l}$ and 0.35 $\mu\text{g}/\text{l}$, respectively. In contrast to fresh surface waters, only limited data are available for sediments and less for marine ecosystems. Many of these studies characterized the sample locations as being downstream of wastewater discharges, receiving waters for industrial facilities, areas susceptible to contamination, urban waterways, or industrial ports.

Measurements of BPA in drinking water and its source waters have been reported in numerous studies by government agencies and other researchers.^{12–15} Several of the studies are described as national monitoring programs. For many of these studies, BPA is one of a long list of analytes, whereas other studies have focused on measurements of BPA only. To date, the available data have not been summarized, analyzed statistically, nor has the relevance to human exposures been assessed.

Globally, the source of drinking water is more or less equally divided between surface water and groundwater (48.22% and 48.23%, respectively), with the balance (3.55%) obtained from desalination of saltwater.¹⁶ In 2006, 54% of the world's population had a piped connection providing drinking water, compared with 33% who used other improved drinking water sources. The remaining 13% of the population relied on unimproved sources.¹⁶

Drinking water treatment typically involves mixing surface water with a coagulant to assist with flocculation of finely divided suspended matter, which may be removed by sedimentation and filtration, and then the filtered water is disinfected by chemical methods, predominantly chlorine-based, or by physical methods such as ultraviolet radiation.¹⁷ Depending upon the surface water,

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additional processes may be used, including activated carbon treatment for the removal of dissolved organic material, demineralization for the reduction of dissolved ions (usually through advanced membrane treatment), and hydrogen sulphide/iron/manganese removal.

Drinking water treatment technologies typically remove 76–99% of the amount of BPA present in source waters.^{18,19} For example, Kleywegt et al.¹⁸ determined that drinking water treatment plants using granulated activated carbon or granulated activated carbon followed by ultraviolet radiation removed 80–99% of the BPA detected in river source water. Stackelberg et al.¹⁹ measured a 76% decrease in BPA from source water to finished water for a drinking water treatment plant consisting of clarification with ferric chloride, primary disinfection with sodium hypochlorite, sand/granulated activated carbon filtration, and secondary disinfection. Snyder et al.¹⁵ evaluated 20 drinking water treatment plants where the frequency of BPA detection (detection limit = 0.005 µg/l) decreased from 44% in source water to 16% in raw water intake, 6% in finished drinking water, and 0% in water within the distribution system. The maximum concentration detected was 0.120 µg/l. Benotti et al.¹² evaluated BPA removal in 19 drinking water treatment plants. BPA was detected in 17% of the surface water-source waters, with a maximum measured concentration of 0.014 µg/l, and the concentration was <0.005 µg/l in all finished waters. BPA was not detected in the groundwater-source water; however, the measured concentration in finished water was 0.025 µg/l.

Human exposure to BPA has been evaluated by characterizing the concentration of BPA in media such as diet, dust, and air.^{9,20–22} Willhite et al.²¹ and AIST⁹ have suggested that ingestion of water is a minor source of exposure compared with food intake. Snyder et al.¹⁵ recently evaluated a relatively small dataset of drinking water samples from the United States and determined a margin of safety of 72,000.

The objectives of this study were to conduct a comprehensive review of BPA measurements in drinking water and source waters (surface water and groundwater), and to determine the relative contributions of drinking water to overall exposure and potential human health risk. The investigation included an exhaustive review of studies reporting monitoring data for BPA in North America, Europe, and Asia. Data for Japan were excluded because a comprehensive review was recently completed,⁹ and the results of the review are compared here. Estimated intakes from drinking water were then compared with overall exposure from all sources, and margin of safety determinations were made using established oral toxicity benchmarks. Finally, BPA exposures were evaluated in the context of recently reported human urinary biomonitoring data. It was beyond the scope of this paper to identify the underlying sources or mechanism of BPA entry into drinking water or source waters (i.e., surface water and groundwater).

METHODS

Identification and Evaluation of Studies

A literature search was conducted to identify environmental monitoring studies published between 1990 and 2010, which reported measurements of BPA in drinking water and its source waters. A two-stage data evaluation process was used similar to Klecka et al.¹¹ Studies were initially scored for completeness in the description of sampling location, date, and procedures, analytical/laboratory methodology, analytical reporting limits, analytical results, quality assurance and quality control sample procedures, and quality assurance/quality control results. Each paper was then carefully reviewed by an analytical chemist with expertise in the analysis of BPA. Studies categorized as “reliable” or “very reliable” in both reviews were used in the subsequent evaluation (see Klecka et al.¹¹ for criteria). There were two studies that lacked sufficient information in English for classification of reliability (i.e., a Norwegian government study²³ and a

study of a Chinese drinking water treatment plant²⁴); therefore, absent information for rejection they were retained for further analysis. All studies are listed in the Supplementary Information available online.

Statistical Treatment of Data

Data from the studies that passed the reliability review described above were then summarized using basic descriptive statistics such as the range, median, and 95th percentile. However, there were three issues that confounded this analysis: depending upon the medium, up to 95% of measured concentrations were reported as less than the detection limit; the detection limits differed between studies by four orders of magnitude; and for some studies, only summary statistics were available limiting our ability to combine studies and perform the statistical analysis. There are a number of available methods to characterize non-detected concentrations.²⁵ In a previous analysis of BPA in surface water and sediment, Klecka et al.¹¹ used the non-parametric Kaplan–Meier method;^{25,26} whereby, distributions of datasets with minimal non-detected concentrations were applied to datasets having limited detected concentrations to enable statistical analysis of data. However, in the present study, BPA was quantified in excess of the detection limit in only 5% of samples of drinking water in North America, and this was considered to be too small a fraction to apply the Kaplan–Meier method. Instead, the distribution of concentrations for each medium is described by grouping or binning the data according to detection limit in the case of samples reported as less than the detection limit, and according to measured concentration, in the case of samples for which concentrations exceed the detection limit. These groupings are used to identify the median and 95th percentile concentrations in addition to the overall minimum and maximum concentrations. One advantage of this

Table 1. Geographic distribution of bisphenol A monitoring data.

Country	Number of studies (number of samples) ^a		
	Drinking water	Surface water-source water	Groundwater-source water
<i>North America</i>			
Canada	3 (130+)	4 (130+)	1 (5)
Mexico	0	0	1 (2)
USA	10 (288+)	14 (612+)	10 (451)
Total	13 (418+)	18 (742+)	12 (458)
<i>Europe</i>			
France	1 (2)	1 (2)	0
Germany	1 (10)	0	0
Italy	1 (6)	1 (8)	1 (2)
Norway	0	1 (12)	0
Spain	1 (7)	7 (46+)	1 (3)
Sweden	1 (34) ^c	0	0
UK	1 (4)	2 (70+)	0
Total	6 (63)	12 (138+)	3 (169) ^b
<i>Asia</i>			
China	9 (25)	4 (80)	0
Iran	1 (1)	0	0
Singapore	1 (1)	0	0
South Korea	0	2 (486)	0
Taiwan	0	1 (120)	0
Total	11 (27)	7 (686)	0
Overall total	30 (508+)	37 (1566+)	15 (627)

^aThe values shown represent the total number of studies followed by the number of samples in parentheses. The actual number of samples is larger than reported here as some studies do not report the number of samples.

^b23 countries are represented.

^cThe data are for raw, not finished drinking water as reported by²⁷ and are not included in subsequent analysis of drinking water.

approach is that it can accommodate many of the studies for which individual data points are not available.

RESULTS

Summary of Environmental Monitoring Studies Utilized in the Analysis

A total of 76 papers or reports, published between 1990 and 2010, were identified that contained data for BPA in North America, Europe, or Asia (excluding Japan) in drinking water and/or source waters. Following the data quality and analytical reviews, 11 papers received a low-reliability ranking or were eliminated because they contained duplicate data, had unreliable analytical methods (e.g., contamination of blanks), or the data were presented in a format that could not be used for further analysis. The 65 papers retained for analysis include 31 papers from North America, 17 from Europe, and 17 from Asia. A summary of each study is presented in the Supplementary Information available online.

Drinking water is divided according to source, that is, surface water, groundwater, or mixed/undefined source. Source waters were identified based on the descriptions provided by the study authors and, therefore, there may be additional data for source waters that could not be identified.

Table 1 summarizes the geographical distribution of available monitoring data listing the number of studies and samples for each country. For some studies, the actual number of samples were not reported; therefore, a “+” is placed after the number of samples to indicate the data are representative of a number of samples greater than reported. Tables 2–4 present summary statistics and the distribution of sample concentrations (detected and not detected) within the pre-defined ranges (e.g., <0.001 µg/l, 0.001 µg/l to <0.01 µg/l, 0.01 µg/l to <0.1 µg/l, 0.1 µg/l to <1 µg/l, and ≥1 µg/l) for drinking water, surface water-source water, and groundwater-source water. As individual data points were not always available, the sum of the number of samples may be less than the total number of samples at the top of the table when individual data points are not available. Details of individual studies are provided in the Supplementary Material available online.

Drinking Water

BPA concentrations in drinking water collected from Canada, USA, six European countries, and three Asian countries are presented in Table 2. The data for finished drinking water (includes effluent from drinking water treatment plants (i.e., finished drinking water ready for distribution) as well as distribution water and tap water) are further categorized according to source as described for each study. The North American data include national monitoring studies of drinking water^{13,15} and two other large studies (54–128 samples).^{12,14} The European and Asian studies of drinking water are all much smaller studies (1–12 samples). Summary statistics for each drinking water study are provided in the Supplementary Information available online.

The detection limits for drinking water vary by four orders of magnitude across all studies, and the frequency of detection was 5%, 52%, and 59%, for North America, Europe, and Asia respectively. The greater detection frequency in Asia and Europe is a function of lower detection limits, not higher concentrations of BPA. The limited detection frequency makes it difficult to compare BPA concentrations between regions, as most of the summary statistics are reported as less than the detection limit. In North America, the median for all sources of drinking water ranged from <0.002 µg/l to <1 µg/l, and the 95th percentile ranged from <0.099 µg/l to <1.6 µg/l. The maximum reported detected concentration of BPA in North American drinking water was 0.099 µg/l. In a few studies with detection limits of 1 µg/l or more

Table 2. Bisphenol A concentrations in drinking water reported for North America, Europe, and Asia.

	North America			Europe			Asia		
	Surface water	Groundwater	Mixed/Unspecified	Surface water	Mixed/Unspecified	Surface water	Surface water	Mixed/Unspecified	
Total number of samples	>171	60	>187	12	17	18	18	9	
Number of samples > detection limit	2	1	19	2	13	15	15	1	
Percent samples > detection limit	<1%	2%	<10%	17%	76%	83%	83%	11%	
Number of samples within concentration range (µg/l)	<DL	<DL	<DL	<DL	<DL	<DL	<DL	<DL	
<0.001	>2	0	0	6	0	1	0	0	
0.001 to <0.01	52	1	19 ^d	0	2	0	2	1	
0.01 to <0.1	0	0	0	0	2	0	10	7	
0.1 to <1	0	0	0	0	0	0	3	0	
1 or >1	116	48	7	4	0	0	0	0	
Concentration (µg/l)									
Minimum	<0.0001	<0.005	<0.002	<0.0002	0.0005	<0.0007	<0.0007	<0.001	
Median	<1	<1	<0.002	<0.0002	<0.002	0.026	0.026	<0.014	
95th percentile	<1.6 (E0.42) ^e	<1	<0.099	<5.1	<0.014	0.19	0.19	<0.097	
Maximum	<1.6 (E0.42) ^e	<1 (E0.2)	<1 (E0.45)	<5.1	0.014	0.317	0.317	<0.097	

Abbreviation: DL, detection limit.

^aThe number of samples quantified at a concentration equal to or exceeding the detection limit.

^bTwo data points with a mean value of 0.0019 µg/l.

^cThe results ranged from 0.0005 µg/l to 0.007 µg/l.

^dIndividual data points were not available to provide breakdown within defined concentration ranges; results ranged from 0.002 µg/l to 0.099 µg/l.

^eE denotes a reported estimated value that is less than detection limit.

Table 3. Bisphenol A concentrations in surface water (identified as a drinking source) reported for North America, Europe, and Asia.

	North America		Europe		Asia	
Total number of samples	>742		>138		686	
Number of samples > detection limit	43		78		586	
Percent samples > detection limit	<6%		<57%		85%	
Number of samples ^a within concentration range ($\mu\text{g/l}$)	<DL	Quantified ^b	<DL	Quantified ^b	<DL	Quantified ^b
<0.001	>2	1	3	2	0	0
0.001 to <0.01	148	40 ^c	26	15	0	501 ^d
0.01 to <0.1	109		15	51	49	
0.1 to <1	0	1	6	9	51	14
1 or >1	440	1	10	1	0	>1 ^e
Concentration ($\mu\text{g/l}$)						
Minimum	<0.0001		<0.0002		0.0022	
Median	<1		<0.006		<0.0155	
95th percentile	<1		<5.1		— ^f	
Maximum	1.9		<5.1 (2.97 ^g)		4.23	

Abbreviation: DL, detection limit.

^aThe individual data points were not available for all studies; therefore, the sum of the number of samples will be less than the total shown above.

^bThe number of samples quantified at a concentration equal to or exceeding the detection limit.

^cIndividual data points were not available to provide breakdown within defined concentration ranges; results ranged from 0.002 $\mu\text{g/l}$ to 0.12 $\mu\text{g/l}$.

^dIndividual data points were not available to provide breakdown within defined concentration ranges; results ranged from 0.0025 $\mu\text{g/l}$ to 0.0965 $\mu\text{g/l}$.

^e71 samples with concentration between 0.037 $\mu\text{g/l}$ and 4.23 $\mu\text{g/l}$; individual data points were not available to provide breakdown.

^fThe 95th percentile could not be calculated as individual data points were not available for all studies.

^gThe maximum detected value.

Table 4. Bisphenol A concentrations in groundwater (identified as a drinking source) reported for North America and Europe.

	North America		Europe	
Total number of samples	458		169	
Number of samples > detection limit	13		67	
Percent samples > detection limit	3%		40%	
Number of samples ^a within concentration range ($\mu\text{g/l}$)	<DL	Quantified ^b	<DL	Quantified ^b
<0.001	0	2	2	0
0.001 to <0.01	28	0	100	>2
0.01 to <0.1	3	6	0	16 ^c
0.1 to <1	3	0	0	
1 or >1	411	5	0	
Concentration ($\mu\text{g/l}$)				
Minimum	0.0004		<0.0002	
Median	<1		<0.001	
95th percentile	<1		<0.073 (90 th percentile)	
Maximum	6.4		2.299	

Abbreviation: DL, detection limit.

^aIndividual data points were not available for all studies; therefore, the sum of the number of samples will be less than the total shown above.

^bThe number of samples quantified at a concentration equal to or exceeding the detection limit.

^cIndividual data points were not available to provide breakdown within defined concentration ranges; results ranged from 0.073 $\mu\text{g/l}$ to 2.299 $\mu\text{g/l}$.

(e.g. Carter *et al.*¹³), BPA was present in several samples and estimates of values less than the detection limit are reported for those samples; the maximum of these estimated concentrations is 0.45 $\mu\text{g/l}$. The predominately lower detection limits in Europe yielded medians ranging from <0.0002 $\mu\text{g/l}$ to <0.002 $\mu\text{g/l}$, and the 95th percentile ranged from <0.014 $\mu\text{g/l}$ to <5.1 $\mu\text{g/l}$ depending on the source of drinking water. Unfortunately, the study of Fawell *et al.*²⁸ with the detection limit of 5.1 $\mu\text{g/l}$ appreciably skews the results. In this study, four samples of finished drinking water from two drinking water treatment plants in the UK were analyzed and BPA was not detected. Similarly for Asia, lower detection limits yielded a median ranging from <0.014 $\mu\text{g/l}$ to 0.026 $\mu\text{g/l}$, and a 95th percentile ranging from <0.097 $\mu\text{g/l}$ to 0.19 $\mu\text{g/l}$ depending on the

source. Asia had the highest BPA concentration quantified (0.317 $\mu\text{g/l}$).

Surface Water Sources

BPA concentrations in surface water-source water collected from Canada, USA, five European countries, and three Asian countries are presented in Table 3. The North American data include national monitoring studies described above for drinking water^{13,15} and two other large studies.^{12,14} In Asia, several large studies of surface water-source water are available (52–480 samples),^{29–31} while the European studies of surface water-source water are smaller in scope (most have 2–12 samples). Summary statistics for each of the studies are provided in the Supplementary Information available online.

Similar to the drinking water data, the detection limits for surface water vary by four orders of magnitude across all studies, and the frequency of detection was 6%, 57%, and 85% for North America, Europe, and Asia respectively. Again, the difference between regions is primarily a function of detection limit sensitivity.

For surface water-source water in North America, the median and 95th percentile concentrations are equal to the predominant detection limit of 1 $\mu\text{g/l}$. For Europe, the median and 95th percentile concentrations are <0.006 $\mu\text{g/l}$ and <5.1 $\mu\text{g/l}$, and the median concentration for Asia is <0.0155 $\mu\text{g/l}$. The maximum concentration quantified in surface water-source water was 4.23 $\mu\text{g/l}$.

Groundwater Sources

BPA concentrations in groundwater-source water collected from Canada, Mexico, USA, and 23 European countries are presented in Table 4. No data are available for Asia. The USA data include two national monitoring studies of source waters.^{13,32} The European dataset includes a study of groundwater-source water in 23 countries.³³ Summary statistics for each of the studies are provided in the Supplementary Information available online.

Again, there was a wide range of detection limits that varied by four orders of magnitude across all studies, and the frequency of detection was 3% and 40% for North America and Europe, respectively. Similar to the other sampled media, the detection limits for the European studies are generally much lower.

For groundwater-source water, the median concentrations are equal to the detection limits; the median concentrations for North America and Europe are <1 and <0.001 $\mu\text{g/l}$, respectively. In North America, the 95th percentile of the concentrations is <1 $\mu\text{g/l}$. In Europe, the 90th percentile is <0.073 $\mu\text{g/l}$ (the 95th percentile could not be calculated because we do not have individual data points for all studies). The maximum concentrations quantified for North America and Europe are 6.4 $\mu\text{g/l}$ and 2.299 $\mu\text{g/l}$, respectively.

DISCUSSION

A systematic evaluation of the data base of BPA drinking water and source water (surface water and groundwater) concentrations was conducted to determine the relative contributions of drinking water to overall exposure and potential human health risk.

An accepted procedure as outlined in Klecka *et al.*¹¹ was used to categorize available studies for acceptability and inclusion into our evaluation. Studies that were included demonstrated standard, validated methodology; however, some studies were designed to detect multiple related analytes in a sample (e.g., several studies were part of a nationwide reconnaissance program), which likely sacrificed some of the analytical sensitivity. Given that the median and 95th percentile values are not detectable for most source categories, it is difficult to compare BPA concentrations across sources and regions. A comparison of the maximum drinking water concentrations across regions and sources, while not ideal, indicates relatively similar values within each source category. A better indicator of the upper-limit concentrations is the 95th percentile value as maximum values may be influenced by single samples and potential outliers. Nonetheless, across regions, maximum drinking water concentrations ranged from 0.014 $\mu\text{g/l}$ to <5.1 $\mu\text{g/l}$, surface water-source water from 1.9 $\mu\text{g/l}$ to <5.1 $\mu\text{g/l}$, and groundwater-source water from 2.299 $\mu\text{g/l}$ to 6.4 $\mu\text{g/l}$ (Tables 2–4). In general, it would be expected that finished drinking water concentrations of BPA would be 10-fold to 100-fold lower than source waters as the efficiency of drinking water treatment plants indicate a removal efficiency in the range of 76–99%.^{18,19}

A recent comprehensive review ($n = 182$) of source water (the source of the water is not specified), finished drinking water, and tap water was conducted in Japan by the Research Center for

Chemical Risk Management.⁹ BPA was detected in 38% of the 74 source water samples, with a maximum concentration of 0.06 $\mu\text{g/l}$ (the most frequent quantitation limit was 0.01 $\mu\text{g/l}$). BPA was detected in 4% of the 74 finished drinking water samples with a maximum measured concentration of 0.01 $\mu\text{g/l}$, and BPA was detected in 8% of the 34 tap water samples with a maximum concentration of 0.007 $\mu\text{g/l}$ (detection limit of 0.003 $\mu\text{g/l}$ and quantitation limit of 0.01 $\mu\text{g/l}$). A recent study of drinking water in Chicago and its source water (Lake Michigan) found that of the 146 samples tested, BPA was detected in 4 drinking water samples, with a maximum concentration of 0.051 $\mu\text{g/l}$, and in 7 source water samples, with a maximum concentration of 0.054 $\mu\text{g/l}$.³⁴

Overall, a vast amount of drinking water and source water data for BPA is available across North America, Europe, and Asia. Our assessment evaluated 65 studies and >2700 samples and AIST⁹ evaluated 182 samples. The entirety of the data indicate that BPA concentrations in treated drinking water are not likely to be greater than about 0.317 $\mu\text{g/l}$ (i.e., the maximum quantified concentration in our study).

Human Exposure to BPA

For the purposes of evaluating human exposure to BPA via drinking water, the highest quantified median and 95th percentile BPA concentration in drinking water (i.e., 0.026 $\mu\text{g/l}$ and 0.19 $\mu\text{g/l}$, detected in Asia) were used to estimate potential drinking water intakes of BPA for children aged 3 months to <6 months, 3 years to <6 years, 6 years to <11 years, and adults. Drinking water intakes were then compared with estimates of overall BPA exposure and recent human biomonitoring data for similar age groups presented by WHO.²² Margin of safety determinations were made using the lowest of available oral toxicity benchmarks (i.e., oral reference doses, tolerable daily intake values, etc). Here, margin of safety is defined as the oral toxicity benchmark divided by the potential exposure. The target margin of safety is ≥ 1 as the associated uncertainty factors (e.g., inter- and intra-species variability) are already included within the derivation of the oral toxicity benchmark.

Oral toxicity benchmarks are available for BPA from several sources and range from 16 $\mu\text{g/kg-bw/day}$ to 50 $\mu\text{g/kg-bw/day}$. The US Environmental Protection Agency's oral reference dose of 50 $\mu\text{g/kg-bw/day}$ is based upon the lowest observed adverse effect (reduced body weight) level of 50 mg/kg/day from a chronic dietary rat study and an uncertainty factor of 1000.³⁵ The European Food Safety Authority's tolerable daily intake of 50 $\mu\text{g/kg-bw/day}$ (which was recently reaffirmed) is based upon a no observed adverse effect level of 5 mg/kg-bw/day established from a multi-generation reproductive toxicity study in rats and an uncertainty factor of 100.³⁶ Health Canada's provisional tolerable daily intake of 25 $\mu\text{g/kg-bw/day}$ is based on a no observed effect level of 25 mg/kg-bw/day from a 90-day rat study and an uncertainty factor of 1000.³⁷ Finally, Willhite *et al.*²¹ recently derived an oral toxicity benchmark of 16 $\mu\text{g/kg-bw/day}$ based on no observed adverse effect levels of 5 mg/kg-bw/day for systemic toxicity in rats and mice and an uncertainty factor of 300.

Potential drinking water intake was determined using the following equation:

$$\begin{aligned} \text{Dose } (\mu\text{g/kg-bw/day}) &= \text{Concentration in drinking water } (\mu\text{g/l}) \\ &\quad \times \text{Ingestion rate (l/day)} \\ &\quad \times \text{Absorption factor (unitless)/Body weight (kg)} \end{aligned}$$

The US Environmental Protection Agency provides mean estimates of drinking water ingestion of 0.56 l/day, 0.38 l/day, 0.51 l/day, and 1.2 l/day for children aged 3 to <6 months, 3 years to <6 years, 6 years to <11 years, and adults, respectively.³⁸ The

body weights for these age groups are 7.4 kg, 18.6 kg, 31.8 kg, and 70 kg³⁸; although 80 kg is cited by US Environmental Protection Agency, the generally accepted value used in risk assessment of 70 kg is used here). Based upon a median concentration of BPA of 0.026 $\mu\text{g/l}$ and 100% absorption,³⁹ the estimated median BPA drinking water intake is 0.0020 $\mu\text{g/kg-bw/day}$, 0.00053 $\mu\text{g/kg-bw/day}$, 0.00042 $\mu\text{g/kg-bw/day}$, and 0.00045 $\mu\text{g/kg-bw/day}$ for children aged 3 to <6 months, 3 years to <6 years, 6 years to <11 years, and adults, respectively. Using the 95th percentile concentration of 0.19 $\mu\text{g/l}$, the corresponding potential BPA drinking water intake values are 0.014 $\mu\text{g/kg-bw/day}$, 0.0039 $\mu\text{g/kg-bw/day}$, 0.0031 $\mu\text{g/kg-bw/day}$, and 0.0033 $\mu\text{g/kg-bw/day}$. The margin of safety for the median water intake compared with the lowest oral toxicity benchmark of 16 $\mu\text{g/kg-bw/day}$ (which includes an uncertainty factor of 300) ranges from 8200 to 38,000 (Table 5). The margin of safety for the 95th percentile ranges from 1100 to 5200 (Table 5). These findings are in agreement with the work of Snyder et al.¹⁵ who determined a margin of safety of 72,000 based upon a maximum detected drinking water concentration of 0.025 $\mu\text{g/l}$ and an oral toxicity benchmark of 50 $\mu\text{g/kg-bw/day}$.

To understand the relative contribution of drinking water to overall exposure, a comparison is made to estimated BPA intakes recently reported by the World Health Organization²² (Table 5). The primary source of BPA exposure is the diet.²² Other sources of exposure such as inhalation of airborne BPA or indirect ingestion of BPA from soil/dust are at least one order of magnitude less than exposure from the diet.²² The WHO presented ranges of dietary intake for children aged 0 to 6 months, 6 months to 3 year, children aged >3 years to adult, and adults, similar to the age range presented above for drinking water. The average of the range reported by WHO²² for each age group is presented in Table 5. For children aged 0 to 6 months and 6 months to 3 years, three potential diets were presented.²² The highest estimated dietary intake of 2.2 $\mu\text{g/kg-bw/day}$ (i.e., represented by the use of polycarbonate bottles and formula only) and 0.55 $\mu\text{g/kg-bw/day}$ (i.e., represented by a diet using polycarbonate bottles and formula and solid food), respectively, was used. Drinking water contributes very little to overall BPA exposure: the 95th percentile drinking water intake (0.0031 $\mu\text{g/kg-bw/day}$ to 0.014 $\mu\text{g/kg-bw/day}$) is only

0.13–0.40% of the 95th percentile total dietary intake for similar age ranges (Table 5). As expected, the margin of safety for total dietary intake is lower compared with drinking water, but acceptable (> 1) using the most conservative oral toxicity benchmark of 16 $\mu\text{g/kg-bw/day}$ that includes an uncertainty factor of 300. The margin of safety for the average dietary intake ranges from 7.2 to 36 and from 4.4 to 13 for the 95th percentile ranges (Table 5).

To estimate dietary exposure to a particular chemical, a sufficiently large database of chemical concentrations in food along with dietary consumption patterns is needed. This often requires an evaluation of numerous studies to determine exposure.²² Biomonitoring data on the other hand provide direct estimates of internal dose that represent all potential sources and routes of exposure. The WHO²² recently evaluated BPA biomonitoring data from North America,^{40,41} Europe,^{42,43} and Southeast Asia.⁴⁴ The WHO reported that average urinary BPA concentrations (representing free and conjugated BPA) were similar across regions and in the range of approximately 1–3.7 $\mu\text{g/l}$.²² They estimated daily exposure by back-calculating from urinary BPA concentrations^{40,42,43,45} by multiplying by age-specific urinary output and dividing by body weight²² giving median exposure estimates of 0.07 $\mu\text{g/kg-bw/day}$, 0.12 $\mu\text{g/kg-bw/day}$, 0.07 $\mu\text{g/kg-bw/day}$ for children aged 1–5 months, 3–5 years, and 6–11 years, respectively, and 0.05 $\mu\text{g/kg-bw/day}$ for the general population aged 6–60+ years. The 95th percentile values were 1.61 $\mu\text{g/kg-bw/day}$, 0.78 $\mu\text{g/kg-bw/day}$, 0.31 $\mu\text{g/kg-bw/day}$, and 0.27 $\mu\text{g/kg-bw/day}$ for the same age groups, respectively.²² These exposure estimates are 2–31-fold lower than the estimates based on dietary exposure. The margin of safety for BPA intake, back-calculated from the biomonitoring data, ranges from 130 to 320 for the median values and 10 to 59 for the 95th percentile values (Table 5). Compared with the intakes calculated from the biomonitoring data, drinking water contributes very little to overall exposure. The 50th percentile drinking water intake is only 0.4–2.8% of the 50th percentile intake based upon biomonitoring data, and the 95th percentile drinking water intake is only 0.5%–1.2% of the 95th percentile intake based upon biomonitoring data, for similar age ranges (Table 5).

Krishnan et al.⁴⁶ recently developed a biomonitoring guidance value for BPA termed the biomonitoring equivalent. The

Table 5. Estimated intakes of bisphenol A and margins of safety.

	Calculated intakes of BPA $\mu\text{g/kg-bw/day}$		Margin of Safety ^a	
	50th percentile	95th percentile	50th percentile	95th percentile
<i>Drinking water intake</i>				
3 to <6 months	0.0020	0.014	8200	1100
3- to <6-year olds	0.00053	0.0039	30,000	4100
6- to 11-year olds	0.00042	0.0031	38,000	5200
Adult	0.00045	0.0033	36,000	4900
<i>Total intake estimated from dietary exposure^b</i>				
0 to 6-months old	2.2 ^c	3.6	7.2	4.4
6-month to 3-year olds	0.55 ^c	2.3	29	7.0
3-year olds to adult	0.45 ^c	1.2	36	13
adults	0.9 ^c	2.6	18	6.2
<i>Total intake estimated from biomonitoring data^b</i>				
1- to 5-month olds	0.07 ^d	1.61 ^d	230	10
3- to 5-year olds	0.12	0.78	130	21
6- to 11-year olds	0.07	0.31	230	52
6- to >60-year olds	0.05	0.27	320	59

^aThe toxicological benchmark of 16 $\mu\text{g/kg-bw/day}$ ²¹ was used for margin of safety (MOS) determinations. The target MOS is 1 as the associated uncertainty factors (e.g., inter- and intra-species variability) are included within the derivation of benchmark.

^bSource, WHO²².

^cThe mean values reported by WHO²² were used.

^dTwo values were given based on differences in urine volume; the greater value is used here.²²

biomonitoring equivalent is defined as "as the concentration or range of concentrations of chemical in a biological medium (blood, urine, or other medium) that is consistent with an existing health-based exposure guidance value such as a reference dose or tolerable daily intake".⁴⁶ The biomonitoring equivalent is derived using chemical-specific pharmacokinetic data to translate the existing toxicological benchmark (used as the basis for the guideline value) to an internal dose in humans.⁴⁷ The advantage of this approach is that the biomonitoring equivalent can be compared directly with the biomonitoring data without having to back-calculate to an exposure dose using urinary output and body weight. For BPA, Krishnan et al.⁴⁶ determined urinary-based biomonitoring equivalent values of 1–2 mg/l derived from Health Canada's provisional tolerable daily intake and US Environmental Protection Agency's reference dose/European Food Safety Authority's tolerable daily intake, respectively. Based upon the urinary BPA concentrations presented by the WHO²² that ranged from <0.45 µg/l to 3.7 µg/l for the median and from 10.13 µg/l to 22.9 µg/l for the 95th percentile and using the lower biomonitoring equivalent value of 1 mg/l, margin of safety values range from 270 to 2200 for median urinary BPA concentrations and from 44 to 99 for the 95th percentile. These values are similar in magnitude to the margin of safety determined by back-calculating exposure as shown above.²²

In conclusion, this study evaluated 65 independent studies and >2700 samples of drinking water and source waters collected in North America, Europe, and Asia. Although high detection limits limited a statistical analysis of the data, this extensive database combined with the data of AIST⁹ indicate that BPA concentrations in treated drinking water are not likely to be >0.317 µg/l (i.e., the maximum concentration quantified in our study). In Asia, where the data were not limited by elevated detection limits, the highest median concentration in drinking water was 0.026 µg/l and the 95th percentile concentration was 0.19 µg/l. A comparison of the calculated intake of BPA via ingestion of drinking water with the intake back-calculated from urinary biomonitoring data shows that drinking water represents 2.8% of the total intake with a margin of safety (using the lowest available oral toxicity benchmark) >1100 for all age groups.

CONFLICT OF INTEREST

The authors are either employed by or have provided consulting services for companies that produce and sell BPA as well as certain other products that contain BPA. KEC and CAS are independent contractors who have worked for government and industry. SGH is employed by ACC and represents the Polycarbonate/BPA Global Group.

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REFERENCES

- 1 ICIS Chemical Business. Chemical profile: bisphenol A. 12 Oct 2008 <http://www.icis.com/Articles/2008/10/13/9162868/chemical-profile-bisphenol-a.html>.
- 2 Chemical Marketing Reporter. *Chemical Profile Bisphenol A*, 20 Dec 2004.
- 3 EC (European Commission). *European Union Complete Risk Assessment including 2003 Risk Assessment Report and 2008 Addendum: 4,4'-Isopropylidenediphenol (Bisphenol-A): CAS No: 80-05-7, EINECS No: 201-245-8, Complete risk assessment in one document*. 2010 <http://esis.jrc.ec.europa.eu/>.
- 4 SRI Consulting. *CEH Product Review Bisphenol A*. July 2010, pp. 7, 11.
- 5 Cousins I.T., Staples C.A., Klečka G.M., and Mackay D. A multimedia assessment of the environmental fate of bisphenol A. *Hum Ecol Risk Assess* 2002; **8**: 1107–1135.
- 6 Staples C.A., Dorn P.B., Klečka G.M., Oblock S.T., and Harris L.R. A review of the environmental fate, effects and exposures of bisphenol A. *Chemosphere* 1998; **36**: 2149–2173.

- 7 Staples C.A., Woodburn K., Caspers N., Hall A.T., and Klečka G.M. A weight of evidence approach to the aquatic hazard assessment of bisphenol A. *Human Ecol Risk Assess* 2002; **8**: 1083–1105.
- 8 Staples C.A., Woodburn K.B., Klečka G.M., Mihaich E.M., Hall A.T., Ortego L., et al. Comparison of four species sensitivity distribution methods to calculate predicted no effect concentrations for bisphenol A. *Human Ecol Risk Assess* 2008; **14**: 455–475.
- 9 AIST (Advanced Industrial Science and Technology). *AIST Risk Assessment Document Series No. 4 Bisphenol A*. Research Center for Chemical Risk Management, Advanced Industrial Science and Technology Tsukuba West: Tsukuba, Ibaraki, Japan, 2007.
- 10 Environment Canada and Health Canada. *Screening Assessment for the Challenge: Phenol, 4,4'-(1-methylethylidene)bis-* (Bisphenol A) *Chemical Abstracts Service Registry Number 80-05-7*. 2008 www.ec.gc.ca/substances/ese/eng/challenge_batch2/batch2_80-05-7.cfm (accessed April 2011).
- 11 Klečka G.M., Staples C.A., Clark K.E., van der Hoeven N., Thomas D.T., and Hentges S.G. Exposure analysis of Bisphenol A in surface water systems in North America and Europe. *Environ Sci Technol* 2009; **43**: 6145–6150.
- 12 Benotti M.J., Trenholm R.A., Vanderford B.J., Holady J.C., Stanford B.D., and Snyder S.A. Pharmaceuticals and endocrine disrupting compounds in U.S. drinking water. *Environ Sci Technol* 2009; **43**: 597–603.
- 13 Carter J.M., Delzer G.C., Kingsbury J.A., and Hopple J.A. Concentration data for anthropogenic organic compounds in ground water, surface water, and finished water of selected community water systems in the United States, 2002–05. *US Geological Survey Data Series* 2007; **268**: 1–30.
- 14 Ontario Ministry of the Environment. *Survey of the Occurrence of Pharmaceuticals and Other Emerging Contaminants in Untreated Source and Finished Drinking Water in Ontario*. Ontario Ministry of the Environment, Toronto, Ontario. PIBS 7269e 2009.
- 15 Snyder S.A., Trenholm R.A., Snyder E.M., Bruce G.M., Bennett E., Pleus R.C., et al. *Toxicological Relevance of EDCs and Pharmaceuticals in Drinking Water*. Awwa Research Foundation, 2008.
- 16 World Water Assessment Programme. *The United Nations World Water Development Report 3: Water in a Changing World*. UNESCO Publishing, and London: Earthscan, Paris, 2009.
- 17 Great Lakes - Upper Mississippi River Board. *10 States Standards - Recommended Standards for Water Works*. Water Supply Committee of the Great Lakes-Upper Mississippi River Board of State and Provincial Public Health and Environmental Managers. Health Research Inc., Health Education Services Division, Albany, NY, 2007.
- 18 Kleywegt S., Pileggi V., Yang P., Hao C., Zhao X., Rocks C., et al. Pharmaceuticals, hormones and bisphenol A in untreated source and finished drinking water in Ontario, Canada — occurrence and treatment efficiency. *Sci Total Environ* 2011; **409**: 1481–1488.
- 19 Stackelberg P.E., Gibs J., Furlong E.T., Meyer M.T., Zaugg S.D., and Lippincott R.L. Efficiency of conventional drinking-water-treatment processes in removal of pharmaceuticals and other organic compounds. *Sci Total Environ* 2007; **377**: 255–272.
- 20 von Goetz N., Wormuth M., Scheringer M., and Hungerbühler K. Bisphenol A: how the most relevant exposure sources contribute to total consumer exposure. *Risk Anal* 2010; **30**: 473–487.
- 21 Willhite C.C., Ball G.L., and McLellan C.J. Derivation of a bisphenol A oral reference dose (RfD) and drinking-water equivalent concentration. *J Toxicol Environ Health, Part B* **11**: 69–146, (2008).
- 22 WHO (World Health Organization). *Joint (FAO/WHO) Expert Meeting to Review Toxicological and Health Aspects of Bisphenol A*. Summary Report. Ottawa, Canada 2010, Food and Agriculture Organization of the United Nations/World Health Organization.
- 23 Fjeld E., Rognerud S., Enge E.K., Borgen A.R., and Dye C. *Indirect Emissions of Pollutants to Lake Mjøsa from Municipal Sewage Plants and Rivers, 2006*. Norsk institutt for vannforskning (NIVA), 5444-2007 2007.
- 24 Shao X., Ma J., and Wen G. Investigation of endocrine disrupting chemicals in a drinking water work located in Songhua River basin. *Huan Jing Ke xue* 2008; **29**: 2723–2728.
- 25 Helsel D.R. More than obvious: better methods for interpreting nondetect data. *Environ Sci Technol* 2005; **39**: 419A–423A.
- 26 Kaplan E.L., and Meier P. Nonparametric estimation from incomplete observations. *J Am Stat Assoc* 1958; **53**: 457–481.
- 27 Hedlund B., Rodhe J., Arner M., Forsberg J., Taaler M., Forsgren A., et al. *Screeningupdrag inom nationell miljöövervakning. Screening av: Bisfenol A*. NV Diariern. 721-1173-03 Mm, WSP 10033045 2006.
- 28 Fawell J.K., Sheahan D., James H.A., Hurst M., and Scott S. Oestrogens and oestrogenic activity in raw and treated water in Severn Trent water. *Water Res* 2001; **35**: 1240–1244.
- 29 Chen T.-C., Shue M.-F., Yeh Y.-L., and Kao T.-J. Bisphenol A occurred in Kao-Pin River and its tributaries in Taiwan. *Environ Monit Assess* 2010; **161**: 135–145.

- 30 Duong C.N., Ra J.S., Cho J., Kim S.D., Choi H.K., Park J.-H., et al. Estrogenic chemicals and estrogenicity in river waters of South Korea and seven Asian countries. *Chemosphere* 2010; **78**: 286–293.
- 31 Ge J., Cong J., Sun Y., Li G., Zhou Z., Qian C., et al. Determination of endocrine disrupting chemicals in surface water and industrial wastewater from Beijing, China. *Bull Environ Contam Toxicol* 2010; **84**: 401–405.
- 32 Focazio M.J., Kolpin D.W., Barnes K.K., Furlong E.T., Meyer M.T., Zaugg S.D., et al. A national reconnaissance for pharmaceuticals and other organic wastewater contaminants in the United States - II) untreated drinking water sources. *Sci Total Environ* 2008; **402**: 201–216.
- 33 Loos R., Locoro G., Comer S., Contini S., Schwesig D., Werres R., et al. Pan-European survey on the occurrence of selected polar organic persistent pollutants in ground water. *Water Res* 2010; **44**: 4115–4126.
- 34 CDWM (Chicago Department of Water Management). *City of Chicago Emerging Contaminant Study Analysis of Endocrine Disrupting Chemicals, Pharmaceuticals, and Personal Care Products*. Chicago, IL, 2011. Available at <http://www.cityofchicago.org/content/dam/city/depts/water/WaterQltyResultsNRpts/analities/Rnd6tbl20110119.pdf>.
- 35 USEPA (United States Environmental Protection Agency). *Integrated Risk Information System (IRIS) Online*. National Center for Environmental Assessment, Cincinnati, OH, 1993.
- 36 EFSA (European Food Safety Authority). EFSA panel on food contact materials, enzymes, flavourings and processing aids (CEF). Scientific opinion on bisphenol A: evaluation of a study investigating its neurodevelopmental toxicity, review of recent scientific literature on its toxicity and advice on the Danish risk assessment of bisphenol A. *EFSA Journal* 2010; **8**: 1829. 110pp.
- 37 Health Canada. *Health Risk Assessment of Bisphenol A from Food Packaging Applications*. Bureau of Chemical Safety, Food Directorate, Health Products and Food Branch, Ottawa, ON, 2008.
- 38 USEPA. *Exposure Factors Handbook: 2011 Edition*. National Center for Environmental Assessment, Washington, DC; EPA/600/R-09/052F. Available from the National Technical Information Service, Springfield, VA, 2011, <http://www.epa.gov/ncea/efh>.
- 39 Dekant W., and Völkel W. Human exposure to bisphenol A by biomonitoring: methods, results and assessment of environmental exposures. *Toxicol Appl Pharm* 2008; **228**: 114–134.
- 40 CDC (Centers for Disease Control and Prevention). *Fourth National Report on Human Exposure to Environmental Chemicals. Updated Tables, July 2010*. National Center for Environmental Health, Division of Laboratory Sciences, Atlanta, GA, 2010.
- 41 Health Canada. *Report on Human Biomonitoring of Environmental Chemicals in Canada, Results of the Canadian Health Measures Survey Cycle 1 (2007–2009)*. Health Canada: Ottawa, Ontario 2010.
- 42 Becker K., Göen T., Seiwert M., Conrad A., Pick-Fuss H., Müller J., et al. GerES IV: phthalate metabolites and bisphenol A in urine of German children. *Int J Hyg Environ Health* 2009; **212**: 685–692.
- 43 Völkel W., Kiranoglu M., and Fromme H. Determination of free and total bisphenol A in urine of infants. *Environ Res* 2011; **111**: 143–148.
- 44 He Y., Miao M., Herrinton L.J., Wu C., Yuan W., Zhou Z., et al. Bisphenol A levels in blood and urine in a Chinese population and the personal factors affecting the levels. *Environ Res* 2009; **109**: 629–633.
- 45 NTP (National Toxicology Program). NTP brief on bisphenol A [CAS No. 80-05-07]. In *NTP-CERHR Monograph on the Potential Human Reproductive and Developmental Effects of Bisphenol A*. Department of Health and Human Services, Research Triangle Park, NC, United States, 2008 pp. 10–64. <http://cerhr.niehs.nih.gov/evals/bisphenol/bisphenol.pdf>.
- 46 Krishnan K., Gagne M., Nong A., Aylward L.L., and Hays S.M. Biomonitoring equivalents for bisphenol A (BPA). *Regul Toxicol Pharmacol* 2010; **58**: 18–24.
- 47 Hays S.M., Aylward L.L., LaKind J.S., Bartels M.J., Barton H.A., Boogaard P.J., et al. Guidelines for the derivation of biomonitoring equivalents: report from the biomonitoring equivalents expert workshop. *Regul Toxicol Pharmacol* 2008; **51**: S4–S15.



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